

Improving the Terrestrial N Cycle Modeling for A Better Estimation of Agricultural NH₃ Emission Under Sustainable Farming Alternatives

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Growing food demand poses a bigger threat to the environment and public health



Can we secure future food supply without sacrificing the clean air?

Intercropping requires less fertilizer to produce the same amount of crops



Intercropping allows crop competition and enhance biological nitrogen fixation



Relative Fertilizer Usage by Province



Assuming all maize and soybean croplands are now adopting maize-soybean intercropping, on average, **42% less fertilizer** is needed to maintain maize yield in each cropland

Fung et al. (in review)

National adoption of intercropping also helps safeguard air quality Such reduced fertilization cuts national agricultural NH_3 emission by 45%

On average, **42% less fertilizer** is needed

Relative NH₃ Emissions (Maize-Soybean)

%



Fung et al. (in review)

Intercropping could be more economic than the current practice in China Avoided Health Cos

Reduced Fertilizer = +US\$0.5b



Avoided Health Costs = +US\$13b



Additional Machinery & Labor = -US\$6.0b



(+93% relative to the current practice)

Net profit = +US\$67b

+



<u>ltem</u>	<u>US\$</u> <u>Per Unit</u>
Maize	0.25/kg
Soybean	0.41/kg
Urea	0.27/kg
Statistical Life	160k
Labor	186.50/ha
Machinery	40.00/ha

Fung et al. (in review)

We will use CESM to evaluate the potential benefits of intercropping under future climate and socioeconomic scenarios



N-cycle in CESM and the missing pathways



We implement into CLM the "multi-stage" NH₃ volatilization scheme from DNDC (Li et al., 2012)

$$\frac{d\left[\mathrm{NH}_{3\,(\mathrm{g})}\right]}{dt} \approx \left[\mathrm{NH}_{4\,(\mathrm{soil})}^{+}\right](1 - f_{\mathrm{ads}})f_{\mathrm{dis}}f_{\mathrm{vol}}\left(\frac{1}{\Delta t}\right)$$



Fraction of soil NH_4^+ adsorbed is determined by an empirical equation for adsorption:

$$f_{ads} = 0.99(7.2733 f_{clay}^3 - 11.22 f_{clay}^2 + 5.7198 f_{clay} + 0.0263)$$

Fraction of dissociated non-adsorbed NH_4^+ : $NH_4^+_{(non-ads)} \rightleftharpoons NH_{3 (aq)} + H^+_{(aq)}$



wind speed (m s⁻



Comparing with a Chinese NH₃ emission inventory (Zhang et al, 2018)

Fung et al. (in prep.)



[CLM5] vs [Zhang2018] F_NH3_VOL_CROP | $\beta_1 = 0.31$ | $R^2 = 0.37$ 0.25 0.20 [CLM5] (Tg) 0.15 China 0.10 0.05 0.00 0.00 0.05 0.10 0.15 0.20 0.25 [Zhang2018] (Tg)

Monthly NH₃ Emission



On-going and Future Work

- Possible reasons for the model-inventory differences:
 - Absence of the canopy reduction factor
 - Inconsistent crop maps
 - Mismatch in fertilization application rates
 - Deviation in prescribed data: soil pH, deposition
- Now: fine-tuning the new NH₃ schemes against field and satellite measurements
- Maria Val Martin is trying to implement:
 - Flux exchange between CLM and CAMchem, including emission of N_2O , $NO_x \& NH_3$ and deposition of NH_4^+
 - Surface dataset of soil pH
- Next: investigating emission scenarios under future climate and their potential feedback mechanisms

Comparing with AMoN site measurement





Figure 2. Yearly averaged surface concentrations (μ g m⁻³, left vertical color bar) from IDAF, AMoN, EMEP and NNDMN data sets plotted on top of the NH₃ IASI satellite column ($\times 10^{16}$ molec cm⁻², right vertical color bar) distribution for 2011 gridded at 0.25° lat $\times 0.5^{\circ}$ long. Columns and relative error (%, bottom left inset) have been calculated as a weighted mean of all IASI measurements within a cell, following equations described in Van Damme et al. (2014a) (columns with an associated relative above 100% have been filtered).

Thank you